JOURNAL OF AGRICULTURAL AND FOOD CHEMISTRY

Degradation of Pentachlorophenol by Potato Polyphenol Oxidase

Mei-Fang Hou,^{+,§} Xiao-Yan Tang,^{§,△} Wei-De Zhang,^{*,†} Lin Liao,[§] and Hong-Fu Wan[§]

⁺School of Chemistry and Chemical Engineering, South China University of Technology, Guangzhou 510641, China

⁹Guangdong Key Laboratory of Agricultural Environment Pollution Integrated Control, Guangdong Institute of Eco-environmental and Soil Sciences, Guangzhou 510650, China

 $^{\Delta}$ Faculty of Forestry and Biotechnology, Zhejiang Agriculture and Forestry University, Lin'an 311300, China

ABSTRACT: In this study, polyphenol oxidase (PPO) was extracted from commercial potatoes. Degradation of pentachlorophenol by potato PPO was investigated. The experimental results show that potato PPO is more active in weak acid than in basic condition and that the optimum pH for the reaction is 5.0. The degradation of pentachlorophenol by potato PPO reaches a maximum at 298 K. After reaction for 1 h, the removal of both pentachlorophenol and total organic carbon is >70% with 6.0 units/mL potato PPO at pH 5.0 and 298 K. Pentachlorophenol can be degraded through dechlorination and ring-opening by potato PPO. The work demonstrates that pentachlorophenol can be effectively eliminated by crude potato PPO.

KEYWORDS: pentachlorophenol, polyphenol oxidase, degradation, enzyme

INTRODUCTION

Pentachlorophenol (PCP, C₆Cl₅OH), an ionizable hydrophobic organic contaminant, has been used extensively in agricultural, industrial, and domestic applications as a component of fungicides, bactericides, herbicides, insecticides, mollus-cides, biocides, and wood preservatives.^{1–3} PCP not only gives rise to accumulation and biological amplification but also causes immunological and endocrine disorders and infertility problems in humans.¹ On the basis of evidence from animal toxicity studies and human clinical data, the U.S. Environmental Protection Agency (U.S. EPA) has classified PCP as a probable human carcinogenic chemical (B2). There are several techniques for the degradation of chlorophenols, such as zero-valent metal-based reduction,^{3,4} advanced oxidation processes (AOPs),⁵⁻¹⁰ and bioremediation.^{11–19} With the aim of producing highly potent nonspecific oxidants, especially to obtain hydroxyl radical (•OH), AOPs have been developed for the degradation of recalcitrant organic contaminants.²⁰ Besides AOPs, enzymes can also pro-duce reactive oxygen species (ROS).^{21–24} Being powerful biocatalysts, enzyme-catalyzed oxidation as the alternative method for the degradation of PCP has attracted much attention in recent years.16-1

Copper-containing enzymes are usually involved in dioxygen binding, activation, and reduction and perform a variety of critical biological functions,²⁵⁻²⁷ during which [•]OH, H₂O₂, and other ROS are produced through the Fenton reaction.²⁸ Polyphenol oxidases (i. e., laccase and tyrosinase) are a group of coppercontaining enzymes that can catalyze the oxidation of phenol derivatives in the presence of O_2 .^{29,30} In the past three decades, polyphenol oxidase (PPO) has been used for the degradation of organic contaminants.^{18,31–33} The substrate specificity and catalytic competence vary greatly for PPOs from different sources.²⁶ Laccase is often found in fungi, and fungal laccase has been widely studied for degrading various contaminants.^{26,34} Compared with the sparse availability and high cost of laccase, plant PPO can be obtained easily and cheaply for potential applications in wastewater and soil treatment.³⁵ However, there are few reports

of the degradation of PCP by plant PPO thus far. Potato is widespread and cheap for extracting PPO,36 and potato PPO is very effective in the removal of dye pollutants.³⁷

The objective of this work is to investigate the degradation of PCP by potato PPO. The degradation mechanism of PCP will be discussed afterward.

MATERIALS AND METHODS

Materials. PCP (98%) was obtained from Aldrich. Acetic acid, isopropanol, ascorbic acid, phosphate, catechol, and other chemicals were purchased in analytical grade from Guangzhou Chemical Co., Guangzhou, China. Potatoes were purchased from the local vegetable market, Guangzhou, China. Methanol and hexane were obtained in HPLC grade from Acros. Unless specified, all chemicals were used as received.

Potato Polyphenol Oxidase. The extraction procedure for the PPO is derived from the literature $^{36-39}$ and modified as follows: Potatoes (100 g) were homogenized with 100 mL of 0.1 M cold phosphate buffer at pH 7.0 in a blender. The homogenate was filtered through cheesecloth and centrifuged at 4500 rpm for 10 min. The supernatant was collected and added into 200 mL of cold acetone under stirring for 15 min before being sealed for 3 h. All steps above were carried out at 277 K. The sediments in the cold acetone solution were collected, centrifuged, and dried with vacuum freezing, and then the potato PPO was obtained and stored at 261 K.

Activity Test. Potato PPO activity was assayed with 0.20 M catechol as a substrate by UV-vis spectrophotometry (TU1800-PC, Beijing, China); 0.20 mL of 0.20 M catechol and 2.7 mL of 50 mM phosphate buffer at pH 6.5 were added with 0.10 mL of 4 mg/mL PPO in a 1 cm light path cuvette. The total assay volume was 3.0 mL. The increase in absorbance at 400 nm and 298 K was recorded automatically. Each sample was assayed in triplicate. One unit of PPO activity was defined as the amount of enzyme that caused a change in absorbance of 0.0010 per min.

| Received: | June 6, 2011 |
|------------|------------------|
| Revised: | October 3, 2011 |
| Accepted: | October 4, 2011 |
| Published: | October 04, 2011 |

Table 1. Pseudo-First-Order Kinetic Constants (k_1) of the Degradation of PCP

| 10 mg/L PCP, 6.0 Units/mL Potato PPO, at 298 K for Initial 30 min | | | | | | | |
|--|-----------|--------|---------|----------|-------|-------|--|
| pН | 4.0 | 4.5 | 5.0 | 6.0 | 7.0 | 8.0 | |
| k_1 (1/min) | 0.019 | 0.025 | 0.045 | 0.022 | 0.025 | 0.014 | |
| R^2 | 0.986 | 0.953 | 0.937 | 0.996 | 0.916 | 0.896 | |
| 10 mg/L PCP, 6.0 Units/mL Potato PPO, at pH 5.0 for Initial 30 min | | | | | | | |
| temperature (| K) 293 | 298 | 3 303 | 308 | 313 | 318 | |
| k_1 (1/min) | 0.0 | 19 0.0 | 45 0.02 | 20 0.017 | 0.012 | 0.004 | |
| R^2 | 0.99 | 93 0.9 | 37 0.90 | 03 0.914 | 0.930 | 0.944 | |
| 10 mg/L PCP, at pH 5.0 and 298 K for Initial 30 min | | | | | | | |
| enzyme dose | (units/mL | .) 1.5 | 3.0 | 4.5 | 6.0 | 7.5 | |
| k_1 (1/min) | • | 0.02 | 0.03 | 2 0.038 | 0.045 | 0.049 | |
| R^2 | | 0.98 | 35 0.98 | 0 0.959 | 0.937 | 0.887 | |
| 6.0 Units/mL Potato PPO, at pH 5.0 and 298 K for Initial 30 min | | | | | | | |
| PCP C_0 (mg/ | /L) 5 | | 10 | 15 | 20 | | |
| k_1 (1/min) | 0. | 058 | 0.045 | 0.019 | 0.008 | | |
| R^2 | 0. | 911 | 0.937 | 0.860 | 0.952 | | |
| | | | | | | | |

Enzymatic Degradation of PCP. Unless specified, batch experiments were carried out for the degradation of PCP by PPO in a 250 mL flask exposed to air at 298 K in the dark. To obtain a high concentration of PCP, ethanol was used with the ratio of 1.0% (v/v) to help the dissolution of PCP in the solution. Experiments for determination of the optimum pH were conducted in 50 mM citrate/phosphate buffer in the range of 3.5-8.0 for the degradation of PCP in the reaction solution of 200 mL. The effects of temperature (293-318 K), the dosage of enzyme (1.5-7.5 units/mL), and the initial concentration of PCP (5-20 mg/L) on its degradation were examined at the optimum pH.

It is reported that the enzymatic oxidation of PPO and the degradation of chlorophenols by AOPs can be described by the pseudo-firstorder kinetics model (eq 1).^{40,41} Thus, the degradation kinetics of PCP by PPO in this work was fitted by the pseudo-first-order kinetic model

$$-\ln\left(\frac{C}{C_0}\right) = k_1 t \tag{1}$$

where *C* and *C*₀ are the concentration of PCP at time *t* and 0 min, respectively, and k_1 is the pseudo-first-order rate constant for the degradation of PCP by PPO (1/min). The plot of $-\ln(C/C_0)$ versus *t* should give a linear relationship from which k_1 can be determined from the slope of the plot.

Characterization and Analysis. The pH value during the reaction was monitored by a pH-meter (pHS-3C, Shanghai, China). The concentration of PCP was determined by UV-vis spectrophotometry (TU1800-PC, Beijing, China) and HPLC (Waters 1525/2487). In HPLC, a mobile phase consisting of 1.0% (w/w) acetic acid (20%) and methanol (80%) was used and PCP was detected by a UV detector at 295 nm.⁴² The total organic carbon (TOC) measurements were measured with a Shimadzu 5000A (Japan). To detect the TOC removal of PCP, several control experiments were conducted in triplicate, and the clear supernatant was taken for analysis. The intermediates were determined by gas chromatography-mass spectrometry (GC-MS, Thermo Trace-DSQ-2000) on a Finnigan Trace DSO Ultra instrument equipped with an Agilent silicon capillary column (0.25 mm inside diameter by 30 m length). The supernatant samples were extracted by hexane and then acetylated by purified acetic anhydride before GC-MS analysis.^{6,8} The efficiency of hexane extraction was no less than 95%.



Figure 1. Effect of pH value on the degradation of 10 mg/L PCP by 6.0 units/mL potato PPO at 298 K for 60 min.

RESULTS AND DISCUSSION

Effect of pH on the Degradation of PCP. It is well-known that changes in the pH value of the reaction solution may not only affect the shape of an enzyme protein but may also change the charge properties of the substrate. The pH value of the reaction solution might have a significant effect on the degradation of PCP by potato PPO. Experiments were carried out for the degradation of 10 mg/L PCP by 6.0 units/mL potato PPO at 298 K. The results are shown in Figure 1, and the pseudo-first-order kinetics constants are listed in Table 1. It can be seen that the optimum pH value is 5.0.

The optimum pH for PPO catalyzing various substrates is different, which might be ascribed to the nature of the sources of enzyme, the properties of substrates or additives, and the purity of enzyme.^{43,44} The transformation of chlorinated hydroxyphenylureas by laccase achieved a maximum between pH 4.0 and 5.0.45 Mushroom PPO catalyzed the transformation of phenol very effectively at pH values ranging from 5.0 to 8.0 with an optimum at pH 7.0.44 The optimum pH values for PPO from Ferula leaves and stems were found to be 7.0 with catechol, 6.5 with (+)-catechin, and 6.0 with 4-methylcatechol, chlorogenic acid, and (-)-epicatechin.⁴⁴ The common pH range for optimal grape PPO activity, as well as other fruits, is known to be pH 5.0-7.0, such as for Victoria grape PPO, with the maximum activity at pH 5.0.30 It can be seen that the optimal pH for the most PPO ranged from 5.0 to 7.0 and that PPO is more active in weak acid conditions than in basic conditions. It is reported that the range between pH 4.0 and 5.0 is better than the lower or higher pH of the reaction solution for the degradation of PCP by horseradish peroxidase in the presence of hydrogen peroxide.¹⁵ The experimental results in this study show that the degradation of PCP by potato PPO achieves a maximum activity at pH 5.0. Thus, a pH value of 5.0 was used in the following experiments.

Effect of Temperature on the Degradation of PCP. There are different optimum temperatures for PPOs catalyzing different substrates such as PPO from *Ferula* leaves at 285 K with catechol, at 286 K with (-)-epicatechin, at 293 K with chlorogenic acid, at 295 K with 4-methylcatechol, and at 303 K with (+)-catechin.⁴⁶

The effect of temperature on the degradation of 10 mg/L PCP by 6.0 units/mL potato PPO was examined at pH 5.0 for 60 min. The results show that the peak of degradation appears at 298 K (Figure 2 and Table 1). When the temperature is below 303 K, the degradation of PCP by potato PPO can be >60% and up to 76% at 298 K. It can be concluded that the effective degradation of PCP can be achieved with cheap potato PPO at room temperature.

Initial Concentration of PCP and PPO Dosage. When the dosage of potato PPO is 6.0 units/mL, the degradation kinetics constants of PCP decrease from 0.058 to 0.008 1/min with the increase of initial concentration of PCP from 5 to 20 mg/L (Table 1). To discover the effect of PPO dosage on the degradation of PCP, experiments were conducted at pH 5.0 and 298 K. The degradation kinetics constants of 10 mg/L PCP rise from 0.022 to 0.049 1/min upon the increase of PPO dosage from 1.5 to 7.5 units/mL (Table 1). It can be seen that the highly



Figure 2. Effect of temperature on the degradation of 10 mg/L PCP by 6.0 units/mL potato PPO at pH 5.0 for 60 min.



Figure 3. TOC removal and degradation of 10 mg/L PCP by 6.0 units/ mL potato PPO at pH 5.0 and 298 K under air or N_2 .

effective degradation of PCP can be achieved with the lower ratio of PPO to PCP.

Mechanism of PCP Degradation. It is reported that the oxidation of phenols by PPO was favored by ensuring an adequate supply O_2 but greatly inhibited by bubbling with N_2 .⁴⁷ Thus, the effect of N_2 on the degradation of PCP by potato PPO was examined. The result shows that there is little degradation of PCP by potato PPO under N_2 (Figure 3), which suggests that O_2 is a key to degrade PCP and there is little adsorption of PCP onto potato PPO. In the experiments, there is also little polymeric substance detected because the dose of PPO and the concentration of PCP are very low.

Figure 3 shows the TOC removal during the degradation of PCP by potato PPO under air. It can be seen that the TOC removal is below 10% within 20 min and increases obviously to 74% after 60 min under air, which suggests the ring-opening degradation of PCP. The PCP removal is 68% within 20 min and increases slowly to 76% under air. It can be deduced that the dechlorination of PCP by potato PPO happened first and rapidly and was then followed by the ring-opening degradation under air.

The reactivity of the copper site to O_2 in the copper-containing enzymes has been discussed in detail.^{27,48–50} The reduction of O_2 by copper-containing enzymes might facilitate the production of ROS through the Fenton reaction.²⁸ The prepared potato PPO as a copper-containing enzyme can be a catalyst to activate O_2 and then produce ROS for degradation of PCP. Isopropanol is an effective scavenger for °OH.^{51,52} When isopropanol was introduced into the reaction solution, the degradation of PCP decreased to near 29%. Thus, °OH may be the important ROS to degrade PCP by potato PPO. Compared with other enzymemediated degradation and AOPs (Table 2), potato PPO is a



Figure 4. Temporal UV-vis spectra of 10 mg/L PCP by 6.0 units/mL potato PPO at pH 5.0 and 298 K for 30 min.

| Table 2. | Comparison a | among AOPs, | Biotechnique, | Potato PPO | , and Other Enz | yme-Mediated De | gradations of PCP |
|----------|--------------|-------------|---------------|------------|-----------------|-----------------|-------------------|
| | | | | | | | |

| treatment process | $C_0^a (mg/L)$ | pH0 ^b | reaction duration | PCP removal (%) | TOC removal (%) | ref |
|------------------------------------|----------------|------------------|-------------------|-----------------|-----------------|-----|
| photo-Fenton | 13.3 | 5.0 | 5 h | 90 | | 8 |
| sonication | 16 | 7.3 | 4 h | 100 | | 9 |
| photodegradation | 10 | | 25 h | 100 | | 10 |
| biodegradation by WRF ^c | 2.7 | | 6 weeks | | 42.4 | 13 |
| $HRP^d + H_2O_2$ | 13.3 | 4-5 | 3 h | 95 | | 15 |
| this work | 10 | 5.0 | 20 min | 68 | 7.2 | |
| | 10 | 5.0 | 1 h | 76 | 74 | |

 ${}^{a}C_{0}$ = initial concentration of PCP (mg/L). b pH₀ = initial pH value of the reaction solution. c WRF, white rot fungi. d HRP, horseradish peroxidase.



Figure 5. Dechlorinated intermediates detected by GC-MS during the degradation of 10 mg/L PCP by 6.0 units/mL potato PPO at pH 5.0 and 298 K within 10 min.

more effective, economical, and environmentally friendly catalyst for degradation of PCP.

The temporal UV—vis spectra of PCP by potato PPO under air are shown in Figure 4. The strong absorbance band at 220 nm can be ascribed to the substituted aromatic ring of PCP. The bands at 250 and 320 nm can be assigned to conjugated system and free electron of O or Cl atoms of PCP. There is an absorption band around 256 nm but no absorption band at 320 nm in the UV—vis absorption of potato PPO solution. During the degradation of PCP by potato PPO, the absorption of PCP at 320 nm decreases sharply within 30 min and the band around 250 nm also decreases, which suggest the dechlorination of PCP by potato PPO. Meanwhile, the adsorption of PCP at 220 nm decreases and has a blue shift, which suggest the ring-opening degradation of PCP by potato PPO.

The degradation of 10 mg/L PCP by 6.0 units/mL potato PPO was conducted at pH 5.0 and 298 K. The intermediates were detected by GC-MS. As shown in Figure 5, the acetylated products of PCP and its dechlorinated intermediates were detected. According to the GC-MS result, tetrachlorohydroquinone (TeCHQ) and tetrachlorocatechol (TeCC) are the primary dechlorinated intermediates during the initial degradation of PCP by potato PPO. High-chlorinated phenols are expected to be more difficult to degrade than low-chlorinated phenols or unsubstituted phenols.^{53,54} Generally, the ortho-Cl and para-Cl of chlorophenols are easily attacked by 'OH or other active species in most AOPs9,55 and microbe- or enzyme-catalyzed oxidation reactions.^{17,56} The dechlorination and the production of more hydroxyl substituent compounds synchronized before the aromatic ring-opening. Thus, PCP can be degraded through dechlorination and ring-opening by potato PPO. On the basis of the results from GC-MS, HPLC, and TOC analysis, it can be seen that the dechlorination is the main step during the initial degradation of PCP. The ring-opening degradation of PCP and the TOC decrease rapidly after the initial dechlorination of PCP by enzyme. As a result, few intermediates were detected by GC-MS analysis.^{33,57,58} Further studies are needed to elucidate the degradation mechanism of PCP by PPO.

In summary, potato polyphenol oxidase was extracted from potatoes and its catalytic activity for the degradation of pentachlorophenol was investigated. The potato polyphenol oxidase shows superb catalytic activity for the degradation of pentachlorophenol in the presence of O_2 at room temperature. This finding paves the way for efficiently eliminating the highly toxic pentachlorophenol in the environment. The work also provides a new insight for the potential application of potato polyphenol oxidase.

AUTHOR INFORMATION

Corresponding Author

*Phone/fax: 86-20-8711 4099. E-mail: zhangwd@scut.edu.cn.

Funding Sources

This work was supported by the National Natural Science Foundation of China (No. 20907011), the Postdoctoral Science Foundation of China (No. 20090450861 and 201003349), and the Ministry of Science and Technology of China (2008AA06Z311).

REFERENCES

(1) Li, M.; Tsai, S. F.; Rosen, S. M.; Wu, R. S.; Bal Reddy, K.; DiCesare, J.; Salamone, S. J. Preparation of pentachlorophenol derivatives and development of a microparticle-based on-site immunoassay for the detection of PCP in soil samples. *J. Agric. Food Chem.* **2001**, *49*, 1287–1292.

(2) Lamprecht, I.; Motzkus, Ch.; Schaarschmidt, B.; Coenen-Stass, D. Pentachlorophenol – an environmental pollutant: microcalorimetric investigations of an ecological model system. *Thermochim. Acta* **1990**, *172*, 87–94.

(3) OAlonso, F.; Beletskaya, I. P.; Yus, M. Metal-mediated reductive hydrodehalogenation of organic halides. *Chem. Rev.* 2002, 102, 4009–4092.

(4) Kim, Y. H.; Carraway, E. R. Dechlorination of pentachlorophenol by zero valent iron and modified zero valent irons. *Environ. Sci. Technol.* **2000**, *34*, 2014–2017.

(5) Christoforidis, K. C.; Louloudi, M.; Deligiannakis, Y. Complete dechlorination of pentachlorophenol by a heterogeneous SiO₂-Feporphyrin catalyst. *Appl. Catal. B: Environ.* **2010**, *95*, 297–302.

(6) Zimbron, J. A.; Reardon, K. F. Fenton's oxidation of pentachlorophenol. *Water Res.* **2009**, *43*, 1831–1840.

(7) Luo, T.; Ai, Z. H.; Zhang, L. Z. Fe@Fe₂O₃ core-shell nanowires as iron reagent. 4. Sono–Fenton degradation of pentachlorophenol and the mechanism analysis. *J. Phys. Chem. C* **2008**, *112*, 8675–8681.

(8) Fukushima, M.; Tatsumi, K. Degradation pathways of pentachlorophenol by photo-Fenton systems in the presence of iron(III), humic acid, and hydrogen peroxide. *Environ. Sci. Technol.* 2001, 35, 1771–1778.

(9) Weavers, L. K.; Malmstadt, N.; Hoffmann, M. R. Kinetics and mechanism of pentachlorophenol degradation by sonication, ozonation, and sonolytic ozonation. *Environ. Sci. Technol.* **2000**, *34*, 1280–1285.

(10) Suegara, J.; Lee, B. D.; Espino, M. P.; Nakai, S.; Hosomi, M. Photodegradation of pentachlorophenol and its degradation pathways predicted using density functional theory. *Chemosphere* **2005**, *61*, 341–346.

(11) Angelinia, V. A.; Orejas, J.; Medina, M. I.; Agostini, E. Scale up of 2,4-dichlorophenol removal from aqueous solutions using *Brassica napus* hairy roots. *J. Hazard. Mater.* **2011**, *185*, 269–274.

(12) Yang, C. F.; Lee, C. M. Pentachlorophenol contaminated groundwater bioremediation using immobilized *Sphingomonas* cells inoculation in the bioreactor system. *J. Hazard. Mater.* **2008**, *152*, 159–165.

(13) Fahr, K.; Wetzstein, H. G.; Grey, R.; Schlosser, D. Degradation of 2,4-dichlorophenol and pentachlorophenol by two brown rot fungi. *FEMS Microbiol. Lett.* **1999**, *175*, 127–132.

(14) Magar, V. S.; Stensel, H. D.; Puhakka, J. A.; Ferguson, J. F. Sequential anaerobic dechlorination of pentachlorophenol: competitive inhibition effects and a kinetic model. *Environ. Sci. Technol.* **1999**, *33*, 1604–1611.

(15) Zhang, G. P.; Nicell, J. A. Treatment of aqueous pentachlorophenol by horseradish peroxidase and hydrogen peroxide. *Water Res.* **2000**, *34*, 1629–1637.

(16) Shim, S. S.; Kawamoto, K. Enzyme production activity of *Phanerochaete chrysosporium* and degradation of pentachlorophenol in a bioreactor. *Water Res.* **2002**, *36*, 4445–4454.

horseradish peroxidase-catalyzed oxidation of pentachlorophenol. *Environ. Sci. Technol.* **1999**, 33, 1408–1412.

(18) Ullah, M. A.; Bedford, C. T.; Evans, C. S. Reactions of pentachlorophenol with laccase from *Coriolus versicolor. Appl. Microbiol. Biotechnol.* **2000**, *53*, 230–234.

(19) Samokyszyn, V. M.; Freeman, J. P.; Maddipati, K. R.; Lloyd, R. V. Peroxidase-catalyzed oxidation of pentachlorophenol. *Chem. Res. Toxicol.* **1995**, *8*, 349–355.

(20) Hoffmann, M. R.; Martin, S. T.; Choi, W.; Bahnemannt, D. W. Environmental applications of semiconductor photocatalysis. *Chem. Rev.* **1995**, *95*, *69–96*.

(21) Schlosser, D.; Höfer, C. Laccase-catalyzed oxidation of Mn^{2+} in the presence of natural Mn^{3+} chelators as a novel source of extracellular H_2O_2 production and its impact on manganese peroxidase. *Appl. Environ. Microbiol.* **2002**, *68*, 3514–3521.

(22) Gonzales, L.; Hernández, J. R.; Perestelo, F.; Carnicero, A.; Falcón, M. A. Relationship between mineralization of synthetic lignins and the generation of hydroxyl radicals by laccase and a low molecular weight substance produced by *Petriellidium fusoideum*. *Enzyme Microb*. *Technol.* **2002**, *30*, 474–481.

(23) Kermer, S. M.; Wood, P. M. Production of Fenton's reagent by cellobiose oxidase from cellulolytic cultures of *Phanerochaete chrysosporium*. *Eur. J. Biochem.* **1992**, 208, 807–814.

(24) Ahuja, D. K.; Bachas, L. G.; Bhattacharyya, D. Modified Fenton reaction for trichlorophenol dechlorination by enzymatically generated H_2O_2 and gluconic acid chelate. *Chemosphere* **2007**, *66*, 2193–2200.

(25) Solomon, E. I.; Chen, P.; Metz, M.; Lee, S. K.; Palmer, A. E. Oxygen binding, activation, and reduction to water by copper-containing enzymes. *Angew. Chem., Int. Ed.* **2001**, *40*, 4570–4590.

(26) Durán, N.; Rosa, M. A.; D'Annibale, A.; Gianfreda, L. Applications of laccases and tyrosinases (phenoloxidases) immobilized on different supports: a review. *Enzyme Microb. Technol.* **2002**, *31*, 907–931.

(27) Lewis, E. A.; Tolman, W. B. Reactivity of dioxygen-copper systems. *Chem. Rev.* 2004, 104, 1047-1076.

(28) Tabner, B. J.; Turnbull, S.; El-Agnaf, O. M. A.; Allsop, D. Direct production of reactive oxygen species from aggregating proteins and peptides implicated in the pathogenesis of neurodegenerative diseases. *Curr. Med. Chem. – Immun., Endocr. Metab. Agents* **2003**, *3*, 299–308.

(29) Burton, S. G. Biocatalysis with polyphenol oxidase: a review. *Catal. Today* **1994**, *22*, 459–487.

(30) Rapeanu, G.; Loey, A. V.; Smout, C.; Hendrickx, M. Biochemical characterization and process stability of polyphenoloxidase extracted from Victoria grape (*Vitis vinifera* ssp. *Sativa*). *Food Chem.* **2006**, *94*, 253–261.

(31) Karam, J.; Nicell, J. A. Potential applications of enzymes in waste treatment. J. Chem. Technol. Biotechnol. **1997**, 69, 141–153.

(32) Couto, S. R.; Herrera, J. L. T. Industrial and biotechnological applications of laccases: a review. *Biotechnol. Adv.* **2006**, *24*, 500–513.

(33) Boolag, J. M.; Chu, H. L.; Rao, M. A.; Gianfreda, L. Enzymatic oxidative transformation of chlorophenol mixtures. *J. Environ. Qual.* **2003**, *32*, 63–69.

(34) Maruyama, T.; Komatsu, C.; Michizoe, J.; Ichinose, H.; Goto, M. Laccase-mediated oxidative degradation of the herbicide dymron. *Biotechnol. Prog.* **2006**, *22*, 426–430.

(35) Durán, N.; Esposito, E. Potential applications of oxidative enzymes and phenoloxidase-like compounds in wastewater and soil treatment. A review. *Appl. Catal. B: Environ.* **2000**, *28*, 83–99.

(36) Alvaro, S. F.; Francisco, L.; Francisco, G. C. Partial purification of soluble potato polyphenol oxidase by partitioning in an aqueous two-phase system. *J. Agric. Food Chem.* **1995**, *41*, 1219–1224.

(37) Khan, A. A.; Husain, Q. Decolorization and removal of textile and non-textile dyes from polluted wastewater and dyeing effluent by using potato (*Solanum tuberosum*) soluble and immobilized polyphenol oxidase. *Bioresour. Technol.* **2007**, *98*, 1012–1019.

(38) Klabunde, T.; Eicken, C.; Sacchettini, J. C.; Krebs, B. Crystal structure of a plant catechol oxidase containing a dicopper center. *Nat. Struct. Biol.* **1998**, *5*, 1084–1090.

(39) Xuan, Y. J.; Endo, Y.; Fujimoto, K. Oxidative degradation of bisphenol A by crude enzyme prepared from potato. *J. Agric. Food Chem.* **2002**, *50*, 6575–6578.

(40) Pera-Titus, M.; García-Molina, V.; Baños, M. A.; Giménez, J.; Esplugas, S. Degradation of chlorophenols by means of advanced oxidation processes: a general review. *Appl. Catal. B: Environ.* **2004**, *47*, 219–256.

(41) Soliva-Fortuny, R. C.; Elez-Martínez, P.; Sebastián-Calderó, M.; Martín-Belloso, O. Kinetics of polyphenol oxidase activity inhibition and browning of avocado purée preserved by combined methods. *J. Food Eng.* **2002**, *55*, 131–137.

(42) Oturan, M. A.; Oturan, N.; Lahitte, C.; Trevin, S. Production of hydroxyl radicals by electrochemically assisted Fenton's reagent: application to the mineralization of an organic micropollutant, pentachlor-ophenol. *J. Electroanal. Chem.* **2001**, *507*, 96–102.

(43) Bayramoğlu, G.; Yakup Arıca, M. Enzymatic removal of phenol and *p*-chlorophenol in enzyme reactor: horseradish peroxidase immobilized on magnetic beads. *J. Hazard. Mater.* **2008**, *156*, 148–155.

(44) Ikehata, K.; Nicell, J. A. Characterization of tyrosinase for the treatment of aqueous phenols. *Bioresour. Technol.* **2000**, *4*, 191–199.

(45) Jolivalt, C.; Neuville, L.; Boyer, F. D.; Kerhoas, L.; Mougin, C. Identification and formation pathway of laccase-mediated oxidation products formed from hydroxyphenylureas. *J. Agric. Food Chem.* **2006**, *54*, 5046–5054.

(46) Erat, M.; Sakiroglu, H.; Kufrevioglu, O. I. Purification and characterization of polyphenol oxidase from *Ferula* sp. *Food Chem.* **2006**, *95*, 503–508.

(47) López-Molina, D.; Hiner, A. N. P.; Tudela, J.; García-Cánovas, F.; Rodríguez-López, J. N. Enzymatic removal of phenols from aqueous solution by artichoke (*Cynara scolymus* L.) extracts. *Enzyme Microb. Technol.* **2003**, 33, 738–742.

(48) Solomon, E. I.; Sundaram, U. M.; Machonkin, T. E. Multicopper oxidases and oxygenases. *Chem. Rev.* **1996**, *96*, 2563–2606.

(49) Solomon, E. I.; Szilagyi, R. K.; George, S. D.; Basumallick, L. Electronic structures of metal sites in proteins and models: contributions to function in blue copper proteins. *Chem. Rev.* **2004**, *104*, 419–458.

(50) Rorabacher, D. B. Electron transfer by copper centers. *Chem. Rev.* **2004**, *104*, 651–698.

(51) Martin, S. T.; Lee, A. T.; Hoffmann, M. R. Chemical mechanism of inorganic oxidants in the TiO₂/UV process: increased rates of degradation of chlorinated hydrocarbons. *Environ. Sci. Technol.* **1995**, 29, 2567–2573.

(52) Ilisz, I.; Dombi, A. Investigation of the photodecomposition of phenol in near-UV-irradiated aqueous TiO_2 suspensions. II. Effect of charge-trapping species on product distribution. *Appl. Catal. A: Gen.* **1999**, *180*, 35–45.

(53) Dabo, P.; Cyr, A.; Laplante, F.; Jean, F.; Ménard, H.; Lessard, J. Electrocatalytic dehydrochlorination of pentachlorophenol to phenol or cyclohexanol. *Environ. Sci. Technol.* **2000**, *34*, 1265–1268.

(54) Ukrainczyk, L.; McBride, M. B. The oxidative dechlorination reaction of 2,4,6-trichlorophenol in dilute suspensions of manganese oxides. *Environ. Toxicol. Chem.* **1993**, *12*, 2005–2014.

(55) Gunlazuardi, J.; Lindu, W. A. Photocatalytic degradation of pentachlorophenol in aqueous solution employing immobilized TiO_2 supported on titanium metal. *J. Photochem. Photobiol. A: Chem.* **2005**, 173, 51–55.

(56) Bae, H. S.; Lee, J. M.; Lee, S. T. Biodegradation of 4-chlorophenol via a hydroquinone pathway by *Arthrobacter ureafaciens* CPR706. *FEMS Microbiol. Lett.* **1996**, 145, 125–129.

(57) Yin, L. F.; Shen, Z. Y.; Niu, J. F.; Chen, J.; Duan, Y. P. Degradation of pentachlorophenol and 2,4-dichlorophenol by sequential visible-light driven photocatalysis and laccase catalysis. *Environ. Sci. Technol.* **2010**, *44*, 9117–9122.

(58) Hammel, K. E.; Tardone, P. J. The oxidative 4-dechlorination of polychlorinated phenols is catalyzed by extracellular fungal lignin peroxidases. *Biochemistry* **1988**, *27*, 6563–6568.